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Capability of Catfish (*Clarias gariepinus*) to Accumulate Hg²⁺ From Water

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Abstract – Mercury is hazardous contaminant that can be accumulated by aquatic organisms such as fishes, mussels etc. Catfish is one of source of animal protein but it also can accumulate Hg²⁺ from water that used in aquaculture. Due to less information about capability of catfish to accumulate Hg²⁺, therefore we studied bioaccumulation of Hg²⁺ that used biokinetic approach (aqueous uptake-rate, and elimination-rate). Nuclear application technique was applied in this study by using radiotracer of ²⁰³Hg. A simple kinetic model was then constructed to predict the bioaccumulation capability of by catfish. The result of experiments were shown that the uptake rate of difference Hg²⁺ concentration were 79.90 to 101.22 ml.g⁻¹.d⁻¹. Strong correlation between uptake rates with increasing Hg²⁺concentration. In addition, the elimination rates were range 0.080 – 0.081 day⁻¹. The biology half time (t_{1/2b}) of Hg²⁺ in whole body catfish were 8.50 – 8.63 days. However, no clear correlation between elimination rate with increasing concentration of Hg²⁺. The calculation of Bio Concentration Factor (BCF) shown catfish have capability to accumulated Hg maximum 1242.69 time than its concentration in water.

Keywords: Radiotracer; Kinetic; Uptake; Elimination; Bioaccumulation

Introduction

Catfish is one of aquaculture species in Indonesia. Catfishes of the genus *Clarias* (Siluroidei, Claridae) are widespread in tropical Africa and Asia (Sudarto, 2007). Because many catfish aquacultures in low quality water, it has possibility contaminated by pollutants such as heavy metals. One of the most hazardous contaminants is mercury (Hg) that input from natural and anthropogenic activities and then will enter. In environment waters, mostly of Hg present as in organic (Hg²+) and organic (methyl mercury) (Wang, 2012; Leermakers *et al.*, 2005). These mercury species have potential to be accumulated in catfish through water and the food web. Consumption of contaminated cat fish by mercury can effect to human health. WHO (2008) explain that some organs such as: the nervous system, the cardiovascular system and the kidneys are the primary targets for toxicity of mercury and mercury compounds. Moreover, the most of sensitive to toxic effect of mercury are developing organ systems (such as the fetal nervous system). Therefore, it is important to find the information of accumulation of Hg²+ in catfish.

A large number bioaccumulation studies of Hg²⁺ and other heavy metals in fresh water fishes have examined but most previous studies measured the concentration of Hg in fish and compare it's concentration in water (Baker et al., 2009; Choy et al., 2009; Casas and Bacher, 2006; Passos et al., 2007; Kasper et al., 2009; Limbong et al., 2004; Eng et al., 1989, Prasetyo, 2009; Riani, 2010; Mustaruddin, 2013). However, these previously studies were not provided information regarding the uptake and elimination kinetics of the Hg²⁺, which are important parameters in interpreting and predicting the Bioaccumulation Factor. On other hand biokinetics of Hg²⁺ in fishes are less well studied. In this study we quantified the biokinetics (uptake and elimination) of Hg²⁺ from water by catfish (*Clarias gariepinus*) to predict its capability to accumulate the Hg² that used radiotracer. A radiotracer technique was used during the present study because it is a very sensitive method and the biokinetics of Hg can be followed non invasively over time (Tsui and Wang, 2004). African catfish (*C. gariepinus*) is one of the popular freshwater fish widely cultured in Indonesia, and is used for human consumption. Herein, we measured a few kinetic parameters (aqueous uptake-rate, and elimination-rate) of Hg²⁺ species in the fish. A simple kinetic model

was then constructed to predict the bioaccumulation capability of Hg²⁺ by catfish from water. This information is important due to both of the environmental assessment in ecosystems and human risk assessment regarding to fish consumption.

Materials and Methods Radiotracer experiments

The bioaccumulation experiment used methods similar to those of Wang and Wong, 2013; Tsui and Wang 2004 with some modifications. The catfish (C. gariepinus) (7.0 to 7.2 cm length) were purchased from a fish farm in Serpong area, South Tangerang, Indonesia. Catfishes were acclimated by maintaining in aerated fresh water and fed with commercial fish feed twice a day. After 14 days acclimatization process, four catfishes were placed in aquarium and exposed to different concentrations of Hg²⁺ (0.001, 0.005 and 0.01 ppm) in the dissolved (0.22-\mum-filtered fresh water) for a total of 12 days. Every Hg²⁺ concentration also spiked ²⁰³Hg²⁺radioisotope into aquarium until each activity concentration was 1 Bq.ml⁻¹. Radiotracer of $^{203}\text{Hg}^{2+}((t_{1/2} = 46.9 \text{ d}, \text{ in } 0.1 \text{ N HCl}, \text{ specific activity} = 3\mu\text{Ci.g}^{-1})$ was produced in Center for Radioisotope and Radiopharmaceutical, BATAN Indonesia. The uptake of radioisotopes was followed non destructively over time. The activity concentration of ²⁰³Hg²⁺ was measured by a gamma spectrometer (NaI detector) at 279 keV, and was corrected for counting efficiency and geometry. At the end of the uptake experiment (8 days), all catfishes were transferred to the uncontaminated running freshwater. The flow rate of the uncontaminated water was set at 1 l.h-1 to avoid recontamination. The elimination of the ²⁰³Hg²⁺ in each catfish is expressed as the percent of the initial activity at the beginning of the elimination experiment. Model

The uptake kinetics were modelled with a single-component first order kinetic model (Whicker and Schultz, 1982, Metian et al., 2011; Wang, 2012, Cardoso et al., 2013)

$$CF_{c} = C_{ss} \left(1 - e^{-R_{c}} \right) \tag{1}$$

 $CF_t = C_{ss} (1 - e^{-k_s})$ (1) where CF_t and CF_{ss} represent activity concentration factor at time t (d) and at steady state (ml.g-1) respectively, and ke represents biological uptake rate constant (d-1). If there was no indication of reaching a steady state during the time of exposure (non-significant fit to model 1), a simple linear regression model was applied. Concentration factor, CF is ratio of activity concentration of ²⁰³Hg in fish tissue to its activity concentration at water.

$$CF_{c} = k_{\omega}t \tag{2}$$

where ku is the slope of regression (uptake in ml. g-1.d-1). Elimination after return to fresh water was modelled using either a single-component exponential model

$$A_{\mathbf{r}} = A_{\mathbf{o}} e^{\mathbf{k}_{\mathbf{c}} \mathbf{r}} \tag{3}$$

where A_0 and A_t are percent of the initial activity at the beginning of the loss experiment and percent at time t of loss experiment. On other had ke is elimination rate constant. The Bio Concentration Factor (BCF) is ratio of the steady-state chemical concentrations in an aquatic water-respiring organism and the water determined in a controlled laboratory experiment in which the test organisms are exposed to a chemical in the water (UK-EPA, 2011)

$$BCF = \frac{k_H}{k_Z} \tag{4}$$

Result and Discussion

According to Wang (2012), speciation of Hg²⁺ is complicated by their binding to various ligands (e.g., chloride and dissolved organic carbon). Furthermore, Wang (2012) explained that differences in Hg speciation may considerably affect its bioavailability and bioaccumulation in aquatic organisms. In this experiment we calculated all biokinetic parameter from dissolved Hg²⁺. On other hand interaction between Hg and dissolved organic matter (DOM) significantly influences the Hg speciation, solubility, mobility and toxicity in aquatic ecosystems (Ci et al., 2011). Using the pore-size filter (0.22 µm) we removed DOM to ensure the mercury speciation was dissolved Hg²⁺.

After 7d experiment, uptake of Hg2+in whole-body catfish displayed linear kinetics and the steady state wasn't reached (Figure 1). The values of Concentration Factor (CF) at the end of experiment were 517.88 to 650.58 ml.g⁻¹. This result indicated that catfish have capability to accumulated Hg²⁺ 517.88 to 650.58 time than it's concentration on water. However these are not representing as value of Bio Concentration Factor (BCF) due to this value have to be determined at steady state condition.

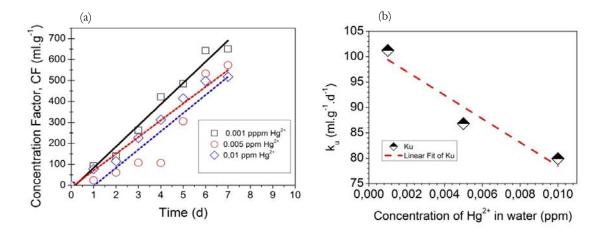


Figure. 1 Bioaccumulation Hg²⁺ (a) Uptake Hg²⁺ from difference concentration (b) Influence of Hg²⁺ concentration in water to uptake rate by catfish

Uptake rate of Hg2+ from difference it's concentration in water were 79.90 to 101.22 ml.g1.d1. Uptake of dissolved metals (Hg2+) from solution through permeable surfaces into the bodies of fish is generally considered to be a passive process (Carvalho et al., 1999). Elevated Hg2+concentration in water induce decreasing of Concentration Factor and uptake rate. This result can be explained based on mechanism bioaccumulation Hg²⁺ through water. Mercury enter to fish via gill that follow the mechanism of respiration or water drinking. According to Morgan et al (2004), it is generally accepted that the key toxic site of action of heavy metal (such as Hg²⁺) is the Na⁺K⁺ATPase located on the basolateral membrane of gill cells. Furthermore, Morgan et al. (2004) explained that this enzyme is responsible for extruding Na⁺in exchange for K⁺ across the basolateral membrane and into the extracellular fluid, thereby providing much of the energy for active Na+ and Cl uptake. In freshwater fish, this transport is essential to counter act the diffusive loss of Na+ and Cl to the hypo osmotic fresh water environment. Water borne Hg²⁺ exposure inhibits the activity of this enzyme causing an inhibition of Na+ and Cl- uptake via the gills (Morgan et al., 2004). Furthermore Stohs and Bagchi (1989) explained that specific differences in the toxicities of metal ions may be related to differences in solubility, absorbability, transport, chemical reactivity, and the complexes that are formed within spite of these factors, the basic mechanisms involving production of reactive oxygen species are the same for these transition metal ions in the body. The toxicity of mercury and its ability to react with and deplete free sulfflydryl groups are well known. Elemental, inorganic, and organic forms of mercury exhibit toxicological characteristics including neurotoxicity, nephrotoxicity, and gastrointestinal toxicity with ulceration and hemorrhage. The decrease in free sulfhydryl groups may lead to the formation of an oxidative stress, resulting in tissue-damaging. (Stohs and Bagchi, 1989)

Regarding the elimination step, this study find that the catfish demonstrated a slow decrease of Hg^{2+} concentration in the first day until the end of the experiment. Probably, the catfish needed more time to progressively detoxify and a continuation to this study would be to assign until achieve equilibrium. When non-contaminating conditions were restored, the whole body elimination kinetics of both Hg^{2+} were best described by a one -component exponential model (Figure 2a). The elimination rate were range 0.080 – 0.081day⁻¹. The biology half time ($t_{1/2b}$) Hg^{2+} in whole body catfish were 8.50 – 8.63 days. However, no clear correlation between elimination rate and increasing concentration of Hg^{2+} because the linear regression coefficient (Adj.R-Sq) bellow 50% (Figure 2b).

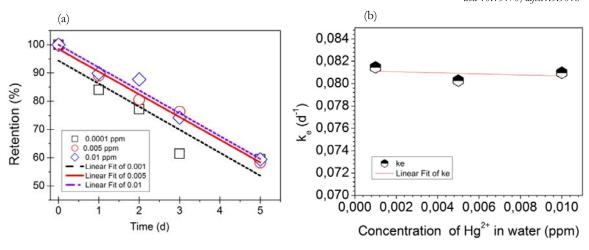


Figure. 2 Elimination of Hg^{2+} (b) Loss of Hg^{2+} in difference concentration, (b) K_e of catfish after accumulated Hg^{2+} in difference concentrations

The uptake experiment only performed for 7 days so that the condition of equilibrium has not been reached. According to Reinardy et al. (2011), dissolved contaminants are primarily taken up across gill membranes or epithelia of the gastrointestinal tract depending on exposure (aqueous or dietary); and, if exposure is of sufficient duration, equilibrium will be established between contaminants in tissues and in the abiotic environment. Another way to define the ability of bioaccumulation is the value of BCF. Van der Oost et al. (2003) explained the bio Concentration Factor (BCF) of a chemical is the ratio of its concentrations in the organism and in water during steady state or equilibrium. The biokinetic parameter that results from this experiment was display on Table 1. The elevation of Hg²⁺ concentration will decrease the uptake rate, elimination rate, Bio Concentration Factor (Table 1). The model bioaccumulation was displayed at Figure 3.

Table 1. Bioakinetic parameter and BCF calculation

Concentration of Hg ²⁺	k_{u}	Ke	BCF	$t_{1/2}$		
in water (ppm)	(ml.g ⁻¹ .day ⁻¹)	(day-1)	$(ml.g^{-1})$	(day)		
0.001	101.22	0.081	1242.69	8.51		
0.005	86.81	0.080	1081.41	8.63		
0.01	79.90	0.080	986.78	8.56		

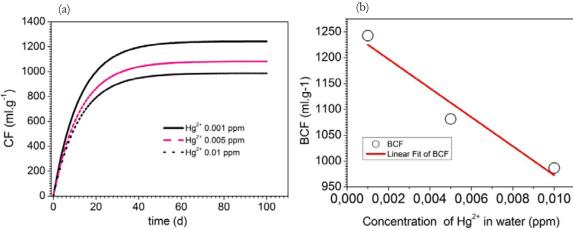


Figure 3. Bioaccumulation model. (a) Prediction steady state condition, (b) Influence concentration Hg²⁺ in water to BCF

Regarding to model, steady state condition was reached after 31 days accumulated Hg²⁺. The BCF have strong correlated with Hg²⁺ concentration in water because increasing concentration will inhibit the metabolism enzyme of catfish. Comparing with another result was shown at Table 2.

Table 2. Biokinetic for fishes using radiotracer techniques

Biota	ku	ke	BCF, calculation	Refference
	$(ml.g^{-1}.day^{-1})$	(day-1)	$(ml.g^{-1})$	
Tilapia (Oreochormis niloticus)	86	0.039	2205.128	Wang et al.
				(2010)
Mosquito fish (Gambusia	52 - 78	0.021 - 0.042	1857.143 - 2476.19	Pickhardt et al.
affanis)				(2006)
Sunfish (Lepomis microlophus)	38 - 51	0.003 -	1033.30 - 1457.14	Pickhardt et al.
		0.0035		(2006)
Catfish (Clarias gariepinus)	79.90 - 101.22	0.080 - 0.081	986.78 – 1242.69	in this study

The result of this experiment was comparable with another Hg²⁺ bioaccumulation experiment that use radiotracer techniques. Furthermore, protection of human health depends directly on the accuracy of estimates of BCF because its variability such as (1) ecological variability (signal) due to ecosystem-specific differences in Hg uptake and accumulation and (2) methodological variability (noise) due to, for example, differences in species, sex, weight, length, age, trophic position, tissue type, collection season, and Hg analysis (Scudder-Eikenberry *et al.*, 2015). Thus, minimizing methodological variability in experiment is critical to BAF-based Hg-risk management. The Hg concentrations in some fishes (including catfish) may be contribute to negative effect to human health, thus Hg exposure to human mainly occurs through dietary intake of contaminated fish (Taylor *et al.*, 2014). Base to this experiment, catfish have capability to accumulated Hg maximal 1242.69 time than its concentration in water. On other hand the threshold levels of Hg²⁺ is 1.0 mg.Kg⁻¹ thus concentration of Hg²⁺ in aquaculture water approximately 0.00081 ppm can give maximum concentration level of Hg in catfish.

Conclusions

Uptake rate of Hg^{2+} from difference it's concentration in water were 79.90 to $101.22 \, \mathrm{ml.g^{-1}.d^{-1}}$. Strong correlation between uptake rates with increasing concentration of Hg^{2+} . The elimination rate were range $0.080 - 0.081 \, \mathrm{day^{-1}}$. The biology half time $(t_{1/2b}) \, Hg^{2+}$ in whole body catfish were $8.50 - 8.63 \, \mathrm{days}$. However, no clear correlation between elimination rate and increasing concentration of Hg^{2+} . Catfish have capability to accumulated Hg (BCF) maximal 1242.69 time than its concentration in water. Due to the threshold levels of Hg in fish products were $1.0 \, \mathrm{mg.Kg^{-1}}$, therefore concentration of Hg^{2+} in aquaculture should be maximum approximately $0.00081 \, \mathrm{ppm}$

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References

- Baker, M., Schindler, D.E., Holtgrieve, G.W. and St Louis, V. 2009. Bioaccumulation and transport of contaminants: migrating sockeye salmon as vectors of mercury. Environmental Science and Technology, 43: 8840–8846.
- Casas, S. and Bacher, C. 2006. Modelling trace metal (Hg and Pb) bioaccumulation in the Mediterranean mussel, *Mytilus galloprovincialis*, applied to environmental monitoring. Journal of Sea Research, 56: 168–181
- Cardoso, P.G., Grilo, T.F., Pereira, E., Duarte, A.C. and Pardal, M.A. 2013. Mercury bioaccumulation and decontamination kinetics in the edible cockle *Cerastoderma edule*. Chemosphere, 90: 1854–1859
- Carvalho, R.A., Benfield, M.C. and Santschi, P.H. 1999. Comparative bioaccumulation studies of colloidally complexed and free-ionic heavymetals in juvenile brown shrimp *Penaeus aztecus* (Crustacea: *Decapoda: Penaeidae*). Limnology Oceanography, 44(2): 403–414.
- Ci,Z., Zhang, X., Wang, Z. and Niu, Z 2011. Phase speciation of mercury (Hg) in coastal water of the Yellow Sea, China. Marine Chemistry, 126: 250–255.

- Choy, C.A., Popp, B.N., Kaneko, J.J and Drazen, J.C. 2009. The influence of depth on mercury levels in pelagic fishes and their prey. Proceedings of the National Academy of Sciences of the United State of America, 106: 13865–13869.
- Eng, C.T., Paw, J.N. and Flordeliz, Y. 1989. Guarin the environmental impact of aquaculture and the effects of pollution on coastal aquaculture development in Southeast Asia. Marine Pollution Bulletin, 20: 335-311.
- Leermakers, M., Baeyens, W., Quevauviller, P. and Horvat, M. 2005. Mercury in environmental samples: speciation, artifacts and validation. Trends in Analytical Chemistry, 24(5): 383-393.
- Kasper, D., Palermo, E.F.A., Dias, A.C.M.I., Ferreira, G.L., Leitão, R.P., Branco, C.W.C. and Malm, O. 2009. Mercury distribution in different tissues and trophic levels of fish from a tropical reservoir, Brazil. Neotropical Ichthyology, 7 (4):751-758.
- Limbong, D., Kamampung, J., Rimper, J. and Arai, T. 2004. Emission and environment implications of mercury from artisanal gold mining related mercury pollution in Rawatotok area of North Sulawesi, Indonesia. Coastal Marine Science, 29 (1):69-74.
- Mustaruddin. 2013. Pola pencemaran Hg dan Pb pada fishing ground dan ikan yang tertangkap nelayan: studi kasus di Teluk Jakarta. Jurnal Bumi Lestari, 13(2):214-224.
- Metian, M., Warnau, M., T eyssié, J.L. and Bustamante, P. 2011. Characterization of ²⁴¹Am and ¹³⁴Cs bioaccumulation in the king scallop *Pectenmaximus*: investigation via three exposure pathways. Journal of Environmental Radioactivity, 102: 543-550.
- Morgan, T.P., Grosell, M., Playle, R.C. and Wood, C.M. 2004. The time course of silver accumulation in rainbow trout during static exposure to silver nitrate: physiological regulation or an artifact of the exposure conditions?. Aquatic Toxicology, 66: 55–72.
- Passos, C.J.S., Mergler, D., Lemire, M., Fillion, M. and Guimarães, J.R.D. 2007. Fish consumption and bioindicators of inorganic mercury exposure. Science of the Total Environment, 373: 68–76.
- Pickhardt, P.C., Stepanova, M.C. and Fisher, N.S. 2006. Contrasting uptake routes and tissue distributions of inorganic and methylmercury in mosquito fish (*Gambusia affinis*) and redear sunfish (*Lepomis microlophus*). Environmental Toxicology Chemistry, 25(8): 2132–2142.
- Prasetyo, A.D. 2006. Penentuan kandungan logam (Hg, Pb Dan Cd) dengan penambahan bahan pengawet dan waktu perendaman yang berbeda pada kerang hijau (*Perna viridis*) di perairan muara kamal, teluk jakarta. Skripsi UIN Syarif Hidayahtullah, Jakarta, 86 pp.
- Riani, E. 2010. Kontaminasi merkuri (Hg) dalam organ tubuh ikan petek (*Leiognathus equulus*) di Perairan Ancol, Teluk Jakarta. Jurnal Teknologi Lingkungan, 11(2): 313-322.
- Reinardy, H.C., Teyssie, J.L., Jeffree, R.A., Copplestone, D., Henry, T.B. and Jha, A. 2011. Uptake, depuration, and radiation dose estimation in zebrafish exposed toradionuclides via aqueous or dietary routes. Science of the Total Environment, 409: 3771–3779
- Scudder-Eikenberry, B.C., Karen Riva-Murray, K.R., Knightes, C.D., Journey, C.A., Chasar, L.C., Brigham, M.E. and Bradley, P.M. 2015. Optimizing fish sampling for fish-mercury bioaccumulation factors. Chemosphere, http://dx.doi.org/10.1016/j.chemosphere. 2014.12.06
- Stohs, S.J and Bagchi, D. 1995. Oxidative mechanisms in the toxicity of metal ions free radical. Biology and Medicine, 18(2): 321-336
- Sudarto, H. 2007. Systematic revision and phylogenetic relationships among populations of clariid species in Southeast Asia. Master Thesis of University of Indonesia, Jakarta, 120 pp.
- Taylor, D.L., Kutil, N.J., Malek, A.J. and Collie, J.S. 2014. Mercury bioaccumulation in cartilaginous fishes from Southern NewEngland coastal waters: Contamination from a trophic ecology and human health perspective. Marine Environmental Research, 99: 20-33.
- Tsui, M.T. and Wang W.X. 2004. Uptake and elimination routes of inorganic mercury and methylmercury in *Daphnia magna*. Environmental Science Technology, 38: 808-816.
- UK-EPA. 2011. Estimation of fish bioconcentration factor (BCF) from depuration data. UK Environment Protection Agency, Horizon House, Deanery Road, Bristol, BS1 5AH.
- Van der Oost, R., Beyer, J. and Vermeulen, N.P.E. 2003. Fish bioaccumulation and biomarkers in environmental risk assessment: a review. Environmental Toxicology and Pharmacology, 13: 57-149
- Wang, X.L. 2012. Biodynamic understanding of mercury accumulationin marine and fresh water fish. Advances in Environmental Research, 1(1): 15-35.
- Wang, R., Wong, M-H. and Wang, W.X. 2010. Mercury exposure in the freshwater tilapia Oreochromis niloticus. Environmental Pollution, 158: 2694-2701.
- Wang, X.L. and Wong, R.S.K. 2003. Bioaccumulation kinetics and exposure pathways of inorganic mercury and methylmercury in amarine fish, the sweetlips *Plectorhinchus gibbosus*. Marine Ecology Progress Series, 261: 257–268.
- Whicker, FW. and Schultz, V. 1982. Radioecology: nuclear energy and the environment. CRC Press, Boca Raton, Fla,185 pp.
- WHO. 2008. Guidance for Identifying Populations at Risk from Mercury Exposure. World Health Organization, Geneva, Switzerland.